Phototreatment of Palm Oil Mill Effluent (POME) over Cu/TiO$_2$ Photocatalyst

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Received: 5th January 2014; Revised: 8th April 2014; Accepted: 8th April 2014

Abstract

The current work reported on the use of different formulations of Cu/TiO$_2$ photocatalysts for the UV-irradiation of palm oil mills effluent (POME). Different copper loadings, viz. 2 wt%, 5 wt%, 10 wt%, 15 wt%, 20 wt% and 25 wt% were doped onto titania. XRD pattern confirmed the presence of anatase TiO$_2$ as primary phase due to mild calcination temperature (573 K). Photo-decomposition of POME over 20 wt% Cu/TiO$_2$ exhibited the highest conversion (27.0%) attributed to its large pore diameter (20.0 nm). In addition, optimum loading was 0.83 g/l. © 2014 BCREC UNDIP. All rights reserved

Keywords: POME; Copper; TiO$_2$; UV-irradiation


Permalink/DOI: http://dx.doi.org/10.9767/bcrec.9.2.6011.121-127

1. Introduction

Malaysia is ranked as one of the largest palm oil producers in the world and generated at least 44 million tones of POME in the year 2008 [1]. Direct discharge of POME to rivers is disastrous to the environment due to its high biochemical oxygen demand (BOD) and chemical oxygen demand (COD) [2]. Therefore, POME normally has to undergo a series of treatments before discharge. However, the conventional open-pond treatment currently used in Malaysia is land-intensive, has long hydraulic retention time (HRT) besides releasing undesirable biogas to the environment [3]. Various treatment and disposal methods have been proposed, investigated and commercialized to reduce POME pollution. However, an ideal treatment process should be cost effective and most importantly, the process should not leave any hazardous residues or at least has to reduce the harmful materials to a safe level.

In the past two decades, photocatalysis process has been touted as one of the most effective clean technology for waste water treatment via removal of organic contaminants [4-6]. During the photoreaction, the catalyst will absorb the UV or visible light-energy from the light-source and attract electrons across the energy gap into the conduction band. This electron will produce hydroxyl radicals which can decompose organic compounds [4]. Hence, the adoption of photocatalysis for POME waste treatment is real especially once the technology has become mature.

According to Fujishima et al. [4], TiO$_2$ is an excellent photocatalyst material for environ-
mental purification. Hence, for the current work, the catalyst used was Cu/TiO$_2$ as this type of catalyst has been claimed as effective [7-9]. Thus, the objective of this study was to undertake systematic analyses into the effects of various photocatalyst formulations towards physicochemical properties and its POME photodegradation ability. Cu of loadings of 2 wt%, 5 wt%, 10 wt%, 15 wt%, 20 wt% and 25 wt% were prepared using wet-impregnation method. Physicochemical properties of the prepared catalysts were determined via several well-established techniques viz. XRD, XRF, BET and TGA. Finally the effectiveness of catalysts were determined by conducting photoreactions. The COD value of POME samples were analysed before and post reaction.

2. Materials and Methods

2.1. Materials

2.1.1. Gases

The gases used in the current study were N$_2$ and He of high purity grade (>99.996%) and were supplied by MOX. Both gases were employed for liquid N$_2$ physisorption analysis.

2.1.2. Chemicals

Table 1 shows the chemicals used. Degussa P25-TiO$_2$ and Cu(NO$_3$)$_2$.3H$_2$O were procured from Sigma-Aldrich. The distilled water was available from the water purification system available in the laboratory of Universiti Malaysia Pahang (UMP) while POME sample was collected from the *Felda Lepar Hilir* 3.

2.1.3. Catalyst Preparation

Degussa P25-TiO$_2$ photocatalyst containing predominantly anatase phase and having a specific surface area of 50 m$^2$/g was sourced from Sigma-Aldrich. The metal dopant precursor, copper nitrate trihydrate, Cu(NO$_3$)$_2$.3H$_2$O (Sigma-Aldrich, >98% purity), was diluted with distilled water to the concentration of 2wt% in a beaker. The accurately-weighed TiO$_2$ was added to the 2 wt% copper nitrate solution that has been prepared in prior and the resulting slurry was magnetic-stirred for 1 h. Subsequently, the slurry-containing beaker was evaporated at 353 K using a water-bath to allow the precursor deposition onto the base TiO$_2$. Finally, the solid material was oven-dried at 393 K for overnight and then air-calcined at 573 K for 30 min. Preparation of other metal loadings, viz. 5 wt%, 10 wt%, 15 wt%, 20 wt% and 25 wt% of copper metal followed the same outlined procedures.

2.3. Catalyst characterization

Catalyst characterization provides useful information on the physicochemical attributes of the catalyst. This section describes the fundamental concepts of various characterization techniques employed for this study. Firstly, X-ray fluorescence (XRF) was employed to determine the actual compositions of the catalysts prepared by using model of S8 Tiger from Bruker Cooperation. Besides, X-ray diffraction was carried out on a Philips X Pert system using CuK$\alpha$ ($\lambda$=1.542 Å) at 30 kV and 15 mA for the identification of structure, crystalline phases, and also sizes of crystallites. Moreover, the density of the prepared catalysts was measured using AccuPyc II 1340 gas pycnometer from Micromeritics. In addition, liquid N$_2$ physisorption analysis was also conducted via Autosorb-1 from Quantachrome Instruments to determine the surface area of the pretreated catalyst in the liquid nitrogen with cross-sectional area of 16.2 Å$^2$ and temperature of 77 K with the help of the Autosorb software, AS1 WIN. This software offers the capability of using either branch of the isotherm for the pore size distribution calculation. Finally, thermogravimetric analysis (TGA) was also performed in an oxidative atmosphere (air) with the ramping rate of 5 K/min for all the prepared catalysts using Q500-model thermogravimetry analyzer (TGA) to determine the thermal stability of the catalysts prepared.

<table>
<thead>
<tr>
<th>Chemical</th>
<th>Purity</th>
<th>Application</th>
</tr>
</thead>
<tbody>
<tr>
<td>Degussa P25-TiO$_2$</td>
<td>80% anatase, 20% rutile</td>
<td>catalyst preparation</td>
</tr>
<tr>
<td>Cu(NO$_3$)$_2$.3H$_2$O</td>
<td>&gt;98%</td>
<td>catalyst preparation</td>
</tr>
<tr>
<td>Distilled water</td>
<td>100%</td>
<td>catalyst preparation, photoreaction</td>
</tr>
<tr>
<td>POME</td>
<td>-</td>
<td>Photoreaction</td>
</tr>
</tbody>
</table>
2.4. Photoreaction

Photoreaction of POME was carried out in a 500-ml pyrex quartz reactor. Transient photodegradation of POME samples was determined via COD analysis. The chemical oxygen demand (COD) of the original POME sample was analyzed using Hach DRB-200 COD reactor before each reaction. Thereafter, 300 ml of POME and 0.25 g of 2 wt% Cu/TiO$_2$ catalyst was mixed in-situ and rigorously stirred for 2 h for adsorption equilibrium. Subsequently, UV lamp (1000 W) was switched on to initiate the reaction. Aliquot 3 ml of POME sample was drawn out from the reactor every 15 min-interval for COD measurements. The COD reduction achieved was recorded. The experiment with catalyst of other metal loadings, viz. 5 wt%, 10 wt%, 15 wt%, 20 wt% and 25 wt% of copper metal followed the same outlined procedures.

3. Results and Discussion

3.1 Catalyst Characterization

The actual percentage of Cu and TiO$_2$ was obtained from XRF analysis and presented in the Table 2. The deviations were ranged from 11% to 22%. The catalyst formulation that has highest deviation was 2 wt% Cu/TiO$_2$ which recorded 22% deviation from the desired wt% while 25 wt% Cu/TiO$_2$ exhibited least deviation from the desired wt% which was only 11%.

From the XRD diffraction pattern, Cu peaks were observed near $2\theta = 12.9^\circ$ and increasing with the desired wt% of Cu, which confirmed the findings of XRF. Besides, CuO crystalline structure was discovered near $2\theta = 36.7^\circ$ with the size ranged from 41.8 to 49.1 nm. Besides, from the results obtained from gas pycnometer, the measured density for all the prepared catalysts have deviated from theoretical density calculated [10]. Figure 1 shows the difference between measured density and theoretical density.

The deviations of the measured density were due to their porous structure. This is consistent with the subsequent findings from liquid N$_2$ physisorption. Besides, liquid N$_2$ physisorption also revealed that the surface area was decreasing with wt% of Cu contained in catalysts. This is due to the formation of CuO crystal upon calcination which has blocked the pores of TiO$_2$ support. In term of specific volume, 5wt% of Cu/TiO$_2$ seems to have the largest adsorption and desorption volume, which were 0.049 and 0.0496 cm$^3$/g respectively. Significantly, the pore volume for both adsorption and desorption decreased in the order of 10 wt% > 15 wt% > 2 wt% > 20 wt% > 25 wt%. In addition, Table 3 presents the pore diameter for the catalysts. It can be seen that the catalysts can be considered as mesoporous as all the pore diameters are fallen into the range of mesoporous material (2-50 nm). The thermal decomposition profile was obtained from TGA study as shown in Figure 2. From Figure 2, one peak was observed for each of the catalysts (besides 2 wt%) at the range of 460 to 540 K symptomatic of the thermal decomposition of copper nitrate to CuO. The derivative weight profile for mass loss area from T= 460 to 540 K were amplified and plotted in Figure 3. Peak for 2 wt% Cu/TiO$_2$ was undetectable probably due to low

Table 2. Results obtained from XRF test

<table>
<thead>
<tr>
<th>Catalyst</th>
<th>Actual wt% of TiO$_2$</th>
<th>Actual wt% of Cu</th>
</tr>
</thead>
<tbody>
<tr>
<td>2 wt% Cu/TiO$_2$</td>
<td>92.79</td>
<td>2.45</td>
</tr>
<tr>
<td>5 wt% Cu/TiO$_2$</td>
<td>88.78</td>
<td>5.65</td>
</tr>
<tr>
<td>10 wt% Cu/TiO$_2$</td>
<td>83.02</td>
<td>11.29</td>
</tr>
<tr>
<td>15 wt% Cu/TiO$_2$</td>
<td>76.77</td>
<td>17.97</td>
</tr>
<tr>
<td>20 wt% Cu/TiO$_2$</td>
<td>70.23</td>
<td>22.88</td>
</tr>
<tr>
<td>25 wt% Cu/TiO$_2$</td>
<td>62.86</td>
<td>27.77</td>
</tr>
</tbody>
</table>

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This was confirmed by the findings of XRD as in the preceding discussion. Based on Figure 3, the maximum peak temperatures seem to be invariant with the Cu loading. However, it can be observed that the intensity of the peaks increased with the wt% of Cu, which indicating higher content of copper nitrate present in the particular catalyst.

### 3.2 Decomposition of POME

Aliquot 3 ml of POME samples was withdrawn every 15 min intervals for 1 h and analysed using COD. The obtained results are presented in Table 4. Over the entire photoreaction, the composition reading from COD showed decreasing trend.

Table 5 shows the conversion for each of the catalysts. The conversion value was computed using Equation 1:

$$ X = \frac{\text{COD}_i - \text{COD}_f}{\text{COD}_i} $$  

where COD$_i$ represents the initial concentration whilst COD$_f$ is the transient conversion. Highest conversion was achieved by 20 wt% Cu/TiO$_2$, which was 27% of the initial organics decomposed within 1 h.

Figure 4 shows the conversion of POME and the XRF results for 2 wt% to 20 wt% of Cu/TiO$_2$. According to the conversion trend, for the Cu content lower than 20%, the conversion was increasing with the Cu content. This can be explained by the higher composition of Cu that has widen the energy band gap of the catalysts for electron excitation and eventually enhancing the reactivity of the catalysts as well as the conversion of the POME samples. However, beyond 20 wt% Cu/TiO$_2$, higher wt% of Cu seems to be non-effective towards reactivity of catalysts. Figure 5 shows the relationship between conversion of POME with pore diameter for 20 wt% and 25 wt% of Cu/TiO$_2$. Based on the results obtained, the conversion profile dropped gradually beyond 20 wt% of Cu/TiO$_2$ and this may be due to pore diameter. As
shown in Table 3, the small desorption pore diameter of 25 wt% Cu/TiO$_2$ has probably inhibit the accessibility of the substrate in POME and eventually limited the catalyst reactivity and POME conversion.

3.3 Kinetic Modelling

The decomposition rate of organics in POME was described by the power-law model given in Equation 2:

$$- \dot{r}_o = kC_A^{\alpha}$$

where (-\dot{r}_o) is the rate of organics decompose (ppm/min), k is the specific reaction rate (ppm/min); $C_A$ is the concentration of organics in POME (ppm) and $\alpha$ is the order of the reaction. The data from Table 4 were analyzed using integral method [11] to determine the reaction order. In the current work, several orders of reaction were tested such as 1$^{st}$, 2$^{nd}$ and 3$^{rd}$. Upon comparison, it was found that the 2$^{nd}$ order fitting seems to give the best fit with the decomposition rate. The results are shown in Figure 6.

The adequacy of power-law model for extracting essential information from organics decomposition data was appraised by a rigorous error analysis. Figure 7 shows a parity plot suggesting a fairly good agreement between predicted ($1/C_A$) value and observed ($1/C_A$) value from Table 4 with $R^2$-value of 0.91. Additionally, as the residual plot depicted in Figure 8 evinces no obvious pattern, hence it can be concluded that power-law model is sufficient [12] to represent the current data.

Moreover, the specific reaction constant (k) obtained were different for each catalysts. The slopes of these data (k) were calculated and tabulated in Table 6. Based on Table 6, these obtained k-values reasonably shows a trend with optimum at 20 wt% of Cu, which was consistent with the Cu composition and pore diameter discussed earlier that lead to higher conversion. The $k$ for 20 wt% Cu/TiO$_2$ (2.6016) was higher than 25 wt% (1.9301), 10 wt% (1.7904), 15 wt% (1.6197), 5 wt% (1.5815) and lastly 2 wt% (1.1913).

3.4 Optimum Catalyst Loading

Three sets of photoreaction were conducted with catalyst loading of 0.5 g/l, 0.83 g/l and 1.0 g/l to determine the optimum catalyst loading for POME photo-degradation. The catalyst cho-
sen for current set of reaction was 20 wt% of Cu/TiO$_2$ as it is previously proven as the optimum performer in terms of conversion. The results obtained are plotted in the Figure 9. Based on the results (cf. Figure 9), the optimum catalyst loading was 0.83 g/l. An approximately 27.0% of the organics was converted within 1 h. Generally, higher catalyst loading indicates higher conversion of organics in POME. Thus, for catalyst loading lower than 0.83 g/l, the conversion of organics would increase with the catalyst loading. Nevertheless, beyond 0.83 g/L, reduction of organics was observed due to high catalyst loading causing formation of slurry mixture which has inhibited the penetration of UV light and eventually causing reduction in organics conversion.

3.5 Longevity Photoreaction Study

Previously, the catalyst with highest reactivity (20 wt% Cu/TiO$_2$) and optimum catalyst loading (0.83 g/l) were determined. A reaction was conducted with these findings to determine the capability of organics decomposition within 7 h. The POME samples were taken for every hour for COD analysis. The conversions were calculated using Equation 1 and results obtained are shown in Figure 10.

Based on COD analysis, after 7 h of UV radiation, more than 40% of organics in POME has been converted. As shown in Figure 10, most of the organics conversion occurred during the 1st h reaction. After that, the reaction progressively slowed down and a final conversion of 40% was achieved after 7 h of UV irradiation. This could be explained by the deactivation of catalyst due to solid organic deposition on the catalyst surface after the organics decomposition.

4. Conclusions

As a conclusion, current work has shown that POME waste can be treated via photoreaction in the presence of Cu/TiO$_2$ catalyst. By using this new proposed method, the retention time for POME treatment can be
significantly decreased from 20 days of hydraulic retention time to within hours. From the results, 20 wt% Cu/TiO₂ yielded the highest organic degradability (conversion) which is 27% of organics converted within 1 h. Besides, the optimum catalyst loading for 20 wt% of Cu/TiO₂ is 0.83 g/l and with this catalyst loading, more than 40% of the organics in POME will decompose within 7 h.

Acknowledgments

The authors would like to thank Ministry of Education Malaysia for the Exploratory Research Grant Scheme RDU120613 and Universiti Malaysia Pahang for Short Term Grant RDU120323.

References


